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## Dependence of ombrotrophic peat nitrogen on phosphorus and climate

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<b>Abstract:</b>	<p>Nitrogen (N) is a key, possibly limiting, nutrient in ombrotrophic peat ecosystems, and enrichment by pollutant N in atmospheric deposition (Ndep, g m<sup>-2</sup> a<sup>-1</sup>) is of concern with regard to peatland damage. We collated data on the N content of surface ombrotrophic peat (Nsp) for 215 sites in the UK and 62 other sites around the world, including boreal, temperate and tropical locations (wider global data), and found Nsp to range from 0.5 % to 4%. We examined the dependences of Nsp on surface peat phosphorus (P) content (Psp), mean annual precipitation (MAP), mean annual temperature (MAT) and Ndep. Linear regression on individual independent variables showed highly significant (<math>p &lt; 0.001</math>) correlations of Nsp with Psp (<math>r^2 = 0.23</math>) and MAP (<math>r^2 = 0.14</math>), and significant (<math>p &lt; 0.01</math>) but weaker correlations with MAT (<math>r^2 = 0.03</math>) and Ndep (<math>r^2 = 0.03</math>). A multiple regression model using log-transformed values explained 36% of the variance of the UK data, 84% of the variance of the wider global data, and 47% of the variance of the combined data, all with high significance (<math>p &lt; 0.001</math>). In all three cases, most of the variance was explained by Psp and MAP, but in view of a positive correlation between MAP and MAT for many of the sites, a role for MAT in controlling Nsp cannot be ruled out. There is little evidence for an effect of Ndep on Nsp. The results point to a key role of P in N fixation, and thereby C fixation, in ombrotrophic peats.</p>	
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**Dependence of ombrotrophic peat nitrogen on phosphorus and climate**

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## ABSTRACT

Nitrogen (N) is a key, possibly limiting, nutrient in ombrotrophic peat ecosystems, and enrichment by pollutant N in atmospheric deposition ( $N_{dep}$ ,  $g\ m^{-2}\ a^{-1}$ ) is of concern with regard to peatland damage. We collated data on the N content of surface (depth  $\leq 25$  cm, mean 15 cm) ombrotrophic peat ( $N_{sp}$ ) for 215 sites in the UK and 62 other sites around the world, including boreal, temperate and tropical locations (wider global data), and found  $N_{sp}$  to range from 0.5 % to 4%. We examined the dependences of  $N_{sp}$  on surface peat phosphorus (P) content ( $P_{sp}$ ), mean annual precipitation (MAP), mean annual temperature (MAT) and  $N_{dep}$ . Linear regression on individual independent variables showed highly significant ( $p < 0.001$ ) correlations of  $N_{sp}$  with  $P_{sp}$  ( $r^2 = 0.23$ ) and MAP ( $r^2 = 0.14$ ), and significant ( $p < 0.01$ ) but weaker correlations with MAT ( $r^2 = 0.03$ ) and  $N_{dep}$  ( $r^2 = 0.03$ ). A multiple regression model using log-transformed values explained 36% of the variance of the UK data, 84% of the variance of the wider global data, and 47% of the variance of the combined data, all with high significance ( $p < 0.001$ ). In all three cases, most of the variance was explained by  $P_{sp}$  and MAP, but in view of a positive correlation between MAP and MAT for many of the sites, a role for MAT in controlling  $N_{sp}$  cannot be ruled out. There is little evidence for an effect of  $N_{dep}$  on  $N_{sp}$ . The results point to a key role of P in N fixation, and thereby C fixation, in ombrotrophic peats.

## Key words:

peat; ombrotrophic; nitrogen; phosphorus; climate; nitrogen deposition; global; nutrient cycling

## INTRODUCTION

The role of nitrogen (N) in peatland ecosystem dynamics has received much recent attention, primarily due to concern about the effects of anthropogenically-driven elevated atmospheric N deposition ( $N_{\text{dep}}$ ) on carbon sequestration (Turunen et al. 2004; Bragazza et al. 2006; Wu et al. 2014;) and biodiversity (Berendse et al. 2001; Chapman et al. 2003; Limpens et al. 2011). Ombrotrophic peats can have a range of N contents; for example in northern peatlands the range is 0.2% to 3% (Loisel et al. 2014). Such variation likely has implications for carbon fixation, peat functioning and sensitivity to increased N inputs.

Peat N contents depend upon inputs from N fixation and atmospheric deposition, and losses by burial into the anaerobic catotelm, leaching, erosion and microbial processing including denitrification. Data compiled by Loisel et al. (2014) imply an average long-term (i.e. thousands of years) burial rate of N in northern peatlands of the order of  $0.5 \text{ g m}^{-2} \text{ a}^{-1}$ . Since  $N_{\text{dep}}$  values of this magnitude are a phenomenon of only the last half-century (Vitousek et al., 1997; Fowler et al., 2004), this accumulation is due almost entirely to inputs by N fixation, which must also account for losses by processes other than burial. Nitrogen fixation rates of the required magnitude, or even greater, have indeed been reported for ombrotrophic bogs (Martin and Holding, 1978; Hemond 1983; Vile et al., 2014).

Whereas N can be acquired by fixation from the atmosphere, P cannot, and this may be significant because P is required for N fixation both as a constituent of the responsible organisms and through the ATP energy-transferring function (Sprent and Raven 1985; Elser et al. 2007; Augusto et al. 2013; Batterman et al. 2013; Reed et al. 2013; Vitousek et al. 2013). Although in most soils, the supply of P is primarily from mineral weathering, this is not the case for ombrotrophic peat, which by definition receives most or all of its inputs from the atmosphere in rain, dust, biological debris from other ecosystems, and the activities of insects, birds and mammals (Rydin and Jeglum 2013; Tipping et al. 2014). Therefore P acquisition is likely a major determinant of variations amongst peats with respect to nutrition, including N fixation. Indeed, the role of P as a limiting factor of biomass growth and functioning in peatlands has been clearly demonstrated (Fritz et al. 2012; Larmola et al. 2013; Hill et al. 2014). Whilst there have been no studies looking specifically at P effects on biological N fixation in peatlands, the findings that P additions increase both peatland plant N uptake (Limpens et al. 2004) and microbial N processing (White and Reddy 2000) illustrate the importance of P in peatland N cycling.

Given the influences of ambient temperature and moisture regimes on biological N cycling (Rustad et al. 2001; Houlton et al. 2008; Ollivier et al. 2011), it is likely that climate also affects N acquisition

by peats. Positive effects of temperature on the N dynamics of peat bogs (Weedon et al. 2012) and on biological N fixation by bryophytic symbionts (Houlton et al. 2008; Lindo et al. 2013) have been demonstrated. Moisture has also been shown to be important for feathermoss-associated N fixation (Gundale et al. 2009; Jackson et al. 2011).

To obtain a wider picture of the possible controlling effects of P and climate on the variation of the N content of ombrotrophic peats, we conducted a meta-analysis of data for a total of 277 sites across boreal, temperate and tropical regions.

## METHODS

We defined three data sets as follows (Table 1): UK-only, wider global (all data except UK), combined (all data). The data were divided between UK and wider global sites because of the much greater number of UK data (see Results). Values for surface peat total N concentrations ( $N_{sp}$ ) and surface peat total P concentrations ( $P_{sp}$ ) measured simultaneously at the same ombrotrophic peatland sites were collated from both published literature and previously unpublished data (Table S1). The previously unpublished data were for UK sites from the Centre for Ecology and Hydrology (A F Harrison pers. comm.), Scottish Soils Database (Hudson *et al.* 2012), and for Finnish sites from the University of Helsinki (R Laiho, pers. comm.). In total our database comprises data from 277 ombrotrophic peatland sites including 215 from the UK, 14 from other temperate localities, 14 from boreal regions and 34 from the tropics (see Table S1 and Figure 1). ‘Surface’ peat was defined as peat sampled from starting depths of 0-10 cm from the surface down to a maximum of 25 cm from the surface. The mean sample depth was 15 cm. Analytical methods for measurements of  $N_{sp}$  and  $P_{sp}$  for each data source are summarised in Table S1. All peat samples had a C concentration  $\geq 40\%$ , the mean C concentration across all sites being 51%. We assume that both N and P in these organic rich soils are overwhelmingly in organic forms. None of the sites considered have been afforested or fertilised. For the UK, however, some sites may have been subjected to variable intensities of drainage.

Values for mean annual precipitation (MAP, m), mean annual temperature (MAT, °C), and total annual N deposition ( $N_{dep}$ ,  $g\ m^{-2}\ a^{-1}$ ) were collated for each site (Table S1). For the UK sites, MAP and MAT are 1970-2000 means from the UK Meteorological Office, and  $N_{dep}$  data are 2006-2008 means derived by the CBED model (Smith *et al.* 2000). For sites not in the UK, MAP and MAT are either values reported in each publication, or 1930-1960 means from the global data set of Cramer and Leemans (2001), with months summed or averaged to give annual values. For all non-UK sites,  $N_{dep}$  data are modelled values for 1993 (Dentener 2006).

## RESULTS

The collated data cover appreciable ranges of  $N_{sp}$ ,  $P_{sp}$ , MAP, MAT and  $N_{dep}$  (Table 1). The values of  $N_{sp}$  vary by a factor of 7 and those of  $P_{sp}$  by a factor of 19, while the NP ratio ranges from 6 to 138. The mapped data (Figure 1) show that the wider global data come from a broad range of locations, although remote peatland localities such as northern Canada and Russia are under-represented. From Table S1 it can be seen that tropical and UK locations have the highest values of  $N_{sp}$ , while NP ratios are lowest for non-UK temperate and boreal sites, and highest for tropical sites, with UK sites in between. Values of MAP and MAT were not significantly correlated for the UK sites, but for the wider global set we found a strong positive correlation which can be parameterised as  $MAP = 0.49 e^{0.077 MAT}$  ( $r^2 = 0.96$ ,  $p < 0.001$ ), and for the combined data set the relationship is  $MAP = 0.93 e^{0.053 MAT}$  ( $r^2 = 0.53$ ,  $p < 0.001$ ). For neither the UK nor the wider global data set was  $N_{dep}$  correlated to MAP or MAT.

Regression analysis of the relationships of  $N_{sp}$  to individual potential driving variables for the combined data set revealed highly significant ( $p < 0.001$ ) positive correlations with  $P_{sp}$  and MAP, and significant ( $p < 0.01$ ) positive correlations with MAT and  $N_{dep}$  (Figure 2). However, none of the relationships explained very much of the variation in  $N_{sp}$  ( $r^2 \leq 0.23$ ). The NP ratio varied positively and significantly with both MAT ( $r^2 = 0.10$ ,  $p < 0.001$ ) and MAP ( $r^2 = 0.11$ ,  $p < 0.001$ ).

Because increased  $N_{dep}$  is a fairly recent phenomenon, and most prevalent in temperate regions, we also conducted a separate analysis of the observations made after 2000 for temperate sites only ( $n = 68$ ). This increased the value of  $r^2$  from 0.03 for the combined dataset ( $n = 277$ ) to 0.07, but the significance was lower ( $p < 0.05$ ). Furthermore, we found that neither UK  $N_{sp}$  nor the UK NP ratio in surface ombrotrophic peat increased with time between 1963 and 2009.

We applied the following multiple regression model to the data;

$$\log N_{sp} = c1 \times \log P_{sp} + c2 \times \log MAP + c3 \times \log (MAT+10) + c4 \times \log N_{dep} + c5 \quad (1)$$

We used log-transformed data to meet the requirements for a normal distribution of the residuals, and added 10 to the MAT values to make them all positive. Because of the imbalance in the spatial distribution of the data, in particular the large number of UK sites, we conducted separate multiple regression analyses of relationships between  $N_{sp}$  and the drivers for UK sites only, wider global data, and combined data. The overall picture was the same in each case, with highly significant dependences on  $P_{sp}$  and MAP and weaker ones on MAT and  $N_{dep}$  (Table 2, Figure 3). Furthermore, the values of the coefficients  $c1$  and  $c2$  were similar for the three data sets, whereas  $c3$  and  $c4$  were variable, and only in two cases are their values significant. The model explained 36%, 84% and 47%

157 of the  $N_{sp}$  variance in the UK, wider global, and combined data sets respectively. The standard errors  
158 in  $\log N_{sp}$  (0.12, 0.09, 0.13) were less diverse than the  $r^2$  values.

159 A simplified model using only  $P_{sp}$  and MAP explained 29%, 84% and 44% of the variances in the UK,  
160 wider global, and combined data sets respectively, with standard errors of 0.12, 0.09 and 0.13 (Table  
161 S2). If MAT was used with  $P_{sp}$ , the fits were poorer although still highly significant ( $p < 0.001$ ),  
162 explaining 27%, 76% and 31% of the variances, with standard errors of 0.12, 0.11 and 0.15 (Table  
163 S3).



## DISCUSSION

The results show that ombrotrophic peat  $N_{sp}$  depends strongly upon  $P_{sp}$  and MAP. The results of the multiple regression analyses are consistent with a multiplicative effect, which can be expressed as;

$$N_{sp} = k P_{sp}^{c1} MAP^{c2} \quad (2)$$

with values of  $k$ ,  $c1$  and  $c2$  of 3.9, 0.35 and 0.44 respectively (Table S2). Because  $c1$  and  $c2$  are both less than one,  $N_{sp}$  is most sensitive to  $P_{sp}$  and MAP when the two drivers have low values, and the relative response decreases as they get larger (Figure S1). The dependence on  $P_{sp}$  is consistent with the need for this element in N fixation (see Introduction), and raises the question as to whether ombrotrophic peats might be P-limited. Indeed P has been found to limit Sphagnum growth at sites receiving high N deposition (Aerts et al. 1992; Gunnarsson and Rydin 2000; Bragazza et al. 2004) and increased investment in P acquisition via phosphatase activity has been observed with peatland N additions (Phuyal *et al.*, 2008).

Both temperature and moisture are likely to affect N accumulation, either through N fixation or by affecting other biogeochemical processes in peats (see Introduction). It also seems possible that the MAP effect arises from seasonal variation, with disruption of N cycling processes occurring during times of moisture deficiency - for example, during periods of low temperature and precipitation in boreal winters and periods of low rainfall in temperate summers. Although significant temperature effects appear when only  $P_{sp}$  and MAT are used as explanatory variables (Table S3), stronger relationships are found with MAP as the second explanatory variable (Table S2), and when both MAP and MAT are included in the multiple regression model, the former is selected as the more explanatory (Table 2). Interpretation here is confounded by the correlation between MAP and MAT, especially in the wider global data set. However, with the UK data set this correlation is not seen, and it may be significant that this is the one instance where both MAP and MAT are significant predictors (Table 2). Therefore we cannot rule out a separate dependence on MAT of  $N_{sp}$ , and it may be that our data are insufficient to draw it out. Nonetheless, it is quite clear that climate exerts a strong effect on the N content of ombrotrophic peats. Furthermore, the positive correlations of peat NP ratio to MAT and MAP suggest that in warmer, wetter regions, proportionally more N is incorporated into surface peat per unit P than in colder, drier regions, which suggests a greater efficiency of P utilisation for N acquisition where climatic conditions favour biological activity.

Our results show that  $N_{sp}$  does not depend strongly on  $N_{dep}$ , even when data are selected to make a fairer comparison by considering only samples collected over a constrained time period, or in a restricted climate zone. There are significant positive responses, but the relationships explain little variation in the data. Although there is evidence that current N deposition influences the N

concentration of *Sphagnum* moss (Bragazza et al. 2005), because elevated  $N_{dep}$  is a recent occurrence, there probably has not been sufficient time for it to affect  $N_{sp}$  as considered here, most of which has instead accumulated via N fixation. Furthermore, it is known that, at least in forest ecosystems,  $N_{dep}$  down-regulates N fixation (DeLuca et al., 2008), and this will tend to cancel any effects of deposition.

The wider global data set is explained very well by equation (2), but not so well the UK data, in terms of  $r^2$ , and this may partly be a statistical artefact because the wider global data are more evenly spread. The SE values (Tables 2, S2, S3) in predicted  $N_{sp}$  are not so different among the three data sets considered, although it is still true that the SE values for the UK-only and combined data sets are higher than that for the wider global set. Whilst to our knowledge the sites included in our analysis were all subject to minimal human disturbance, current and past management practices such as drainage, grazing and burning may have affected their nutrient status (Ramschunder et al. 2009; Jauhiainen et al. 2012; Andersen et al. 2013). This is particularly the case for the UK with its long history of upland management for livestock and grouse rearing (JNCC 2011), and site specific variations in land-management practices may therefore have contributed to the weaker correlation between surface peat N and surface peat P concentrations and climate for the UK sites. Other factors which might account for the unexplained variance in the data include plant type, the effects of atmospherically-deposited contaminants (sulphur, heavy metals, persistent organic pollutants), and the availability of other nutrients.

The great current interest in the role of peatlands in regional and global carbon cycles has resulted in the publication of major reviews (e.g. Limpens et al. 2008; Lindsay 2010; Yu 2012), and the development of sophisticated models (Frolking et al. 2010; Heinemeyer et al. 2010), but only recently has attention has been focused on the role of nutrients and nutrient stoichiometry in carbon fixation (Wu and Blodau 2013; Wang et al. 2014, 2015). As noted by Vile et al. (2014), ombrotrophic peats are highly efficient at fixing C, having net primary production values typically of several hundred  $g\ m^{-2}\ a^{-1}$  despite their low nutrient status. This is due to the low nutrient contents of their vegetation and high nutrient use efficiency (Small 1972; Wang et al 2014).

However, accumulating peats have to combat the loss of nutrients by burial in the catotelm, and while peatland plants may actively hold nutrients in the top layers of peat bogs (Malmer 1998) perhaps by mycorrhizal uptake (Wang et al. 2014), they still bury a good deal of N (Loisel et al. 2014), which necessitates high rates of N fixation. Indeed, the N fixation rates of 1 to 3  $gN\ m^{-2}\ a^{-1}$  reported for bogs by Martin and Holding (1978), Hemond (1983) and Vile et al. (2014) are comparable to the highest rates estimated for different global ecosystems by Cleveland et al. (1999). Our results

strongly suggest that a key factor in the ability of peatlands to carry out N fixation, and thereby C fixation, is P availability, with important modification by climatic conditions, especially precipitation. It seems especially important to understand how peatlands, especially remote ones, acquire P, and how this may have varied over time, given for example Holocene-scale variations in dust transfer (Cockerton et al. 2014) and recent anthropogenic enhancement of this flux (Neff et al. 2008). The incorporation of N and P cycling into models of peat growth is a pressing need.

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## SUPPLEMENTARY INFORMATION

Table S1	Surface peat N and P database
Table S2	Multiple regression analysis results $P_{sp}$ and MAP only.
Table S3	Multiple regression analysis results $P_{sp}$ and MAT only.
Figure S1	Predicted dependence of $N_{sp}$ on $P_{sp}$ at different MAP values ( $m\ yr^{-1}$ )

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471 Table 1. Summary of data. See Table S1 for details.

Data source	n	time period	N <sub>sp</sub> %	P <sub>sp</sub> %	N:P	MAP m	MAT °C	N <sub>dep</sub> g m <sup>-2</sup> a <sup>-1</sup>
UK <sup>1</sup>	215	1963-2009	0.5 - 3.6	0.01 - 0.19	11 - 138	0.8 - 2.8	2.7 - 10.8	0.4 - 3.0
Wider global <sup>2</sup>	62	1971-2012	0.5 - 2.9	0.02 - 0.15	6 - 85	0.4 - 4.0	-3.8 - 26.4	0.0 - 1.9
Combined	277	1963-2012	0.5 - 3.6	0.01 - 0.19	6 - 138	0.4 - 4.0	-3.8 - 26.4	0.0 - 3.0

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473 <sup>1</sup> From: Scottish Soils Database; Emmett et al. 2007; Tipping et al. 2003; AF Harrison pers commun;  
474 Hayati and Proctor, 1991

475 <sup>2</sup> From: Minkinen et al. 1999; Moore et al. 2008; Bragazza et al. 2005; Turetsky et al. 2000;  
476 Pakarinen and Gorham 1984; Richardson et al. 1978; Damman 1978; R Laiho pers commun; Keller et  
477 al. 2006; Bragazza and Gerdol 1999; Bragazza and Gerdol 2002; Clarkson et al. 2004a; Clarkson et al.  
478 2004b; Bridgham et al. 1998; Hill et al. 2014; Cheesman et al. 2012; Page et al. 1999; Anderson 1983;  
479 Pajunen 1994

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Table 2. Multiple regression analysis results; dependence of  $N_{sp}$  on  $P_{sp}$ , MAP, (MAT+10) and  $N_{dep}$  for log-transformed data; coefficients c1-c5 refer to equation (1)

	Variable Coefficient	$P_{sp}$ c1	MAP c2	MAT+10 c3	$N_{dep}$ c4	const c5	$r^2$	SE	p
UK	Value	0.30	0.35	0.89	-0.09	-0.52	0.36	0.12	$1.6 \times 10^{-19}$
	SE	0.04	0.07	0.21	0.05	0.27			
	p	$1.2 \times 10^{-10}$	$2.9 \times 10^{-7}$	$4.4 \times 10^{-5}$	0.057	0.055			
Wider global	Value	0.43	0.56	0.03	-0.02	0.51	0.84	0.09	$8.2 \times 10^{-22}$
	SE	0.07	0.11	0.17	0.03	0.23			
	p	$8.2 \times 10^{-8}$	$2.2 \times 10^{-6}$	0.87	0.51	0.032			
Combined	Value	0.33	0.37	0.11	0.09	0.44	0.47	0.13	$6.2 \times 10^{-37}$
	SE	0.04	0.06	0.10	0.02	0.12			
	p	$2.0 \times 10^{-14}$	$4.2 \times 10^{-9}$	0.30	$6.0 \times 10^{-5}$	$3.1 \times 10^{-4}$			

**Figure captions**

**Figure 1.** Map showing ombrotrophic peat sites. The numbers insider the symbols are the numbers of data for each country or region.

**Figure 2.** Relationships between surface peat %N ( $N_{sp}$ ) and surface peat %P ( $P_{sp}$ ), mean annual precipitation (MAP), mean annual temperature (MAT) and atmospheric N deposition ( $N_{dep}$ ) for the combined dataset. Trend lines and  $r^2$  are for linear regression ( $n = 277$ ); the regressions are all significant, %P and MAP both  $p < 0.001$ , MAT and  $N_{dep}$  both  $p < 0.01$ .

**Figure 3.** Observed  $N_{sp}$  vs. values predicted from linear multiple regressions with  $P_{sp}$ , MAT, MAP and  $N_{dep}$  as independent variables. The 1:1 lines are shown. Numbers of data points are given in Table 1.



501

502 Figure 1

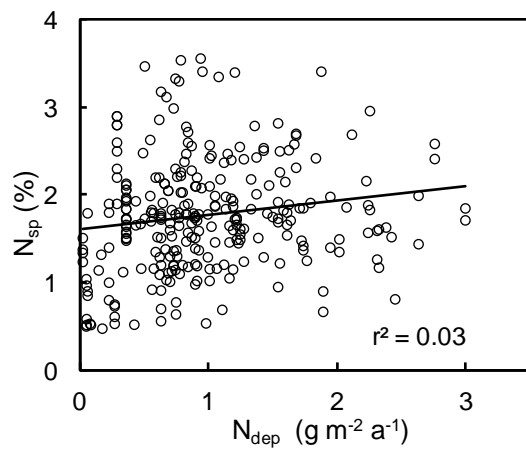
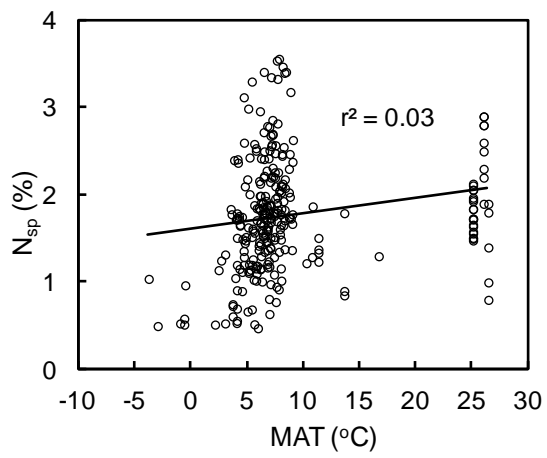
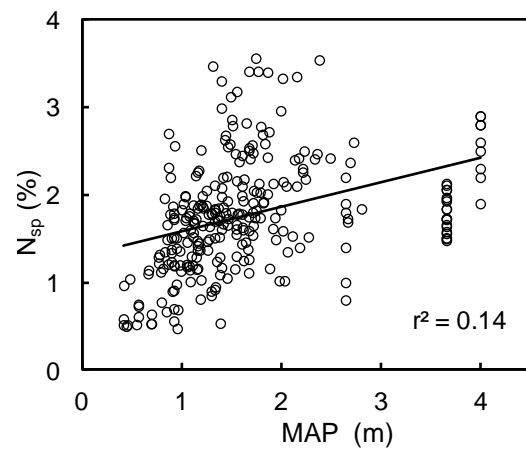
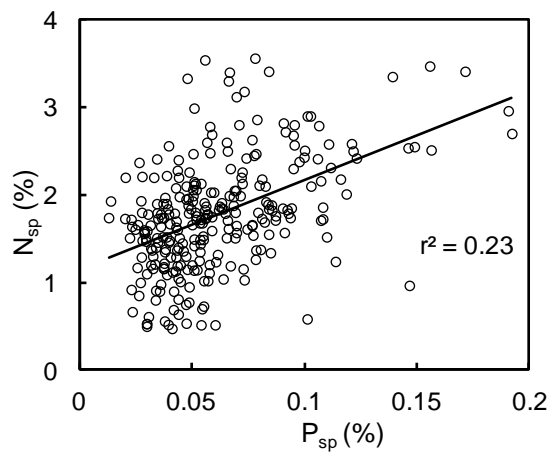


Figure 2

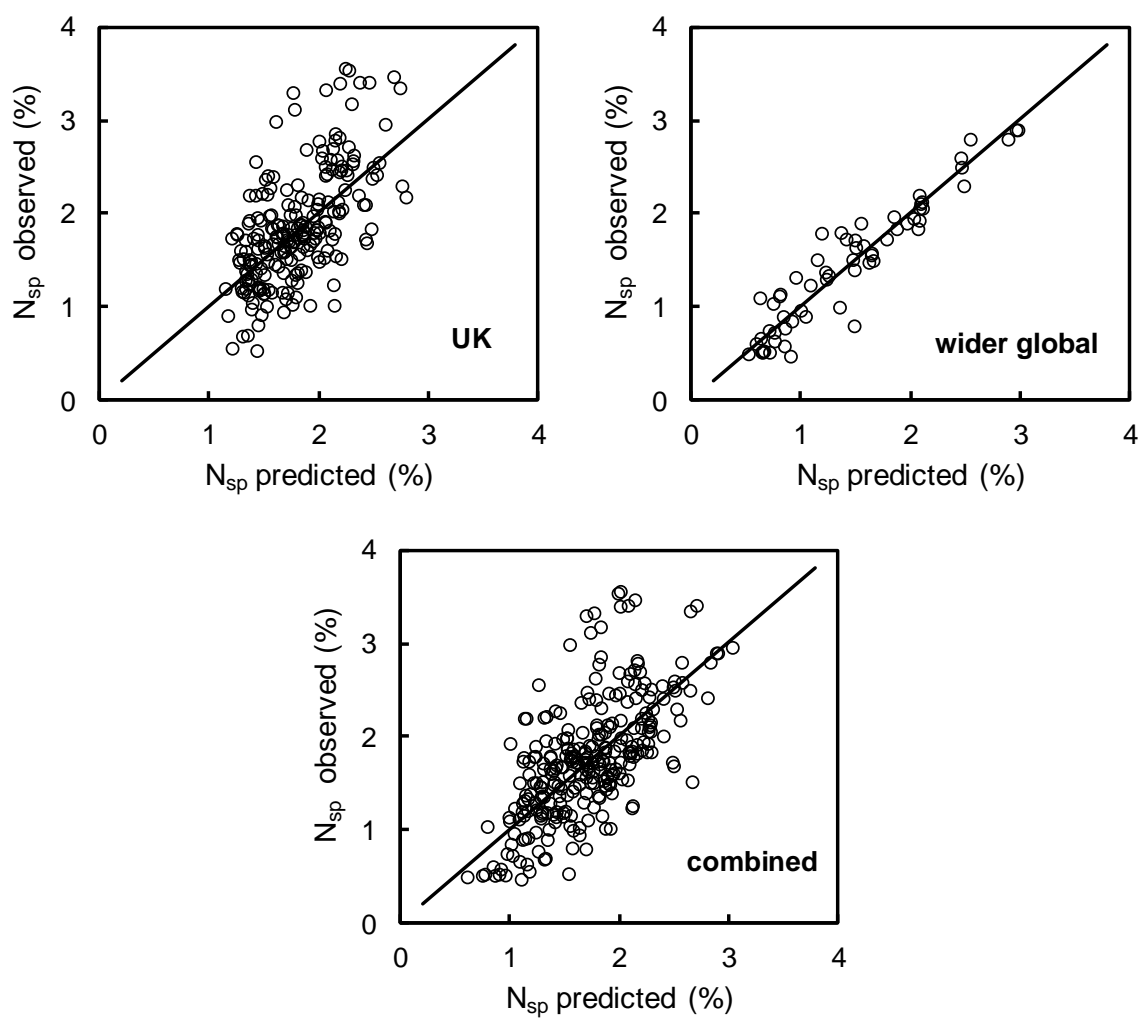


Figure 3



## RESPONSES TO REVIEWERS

Please note that line numbers in the responses refer to the revised text.

### Reviewer #1:

This manuscript presents evidence that the surface concentration of N in ombrotrophic bogs is influenced by P concentration and mean annual precipitation, based on data collected from sites in the UK and a broader data base. The relationship to climate is 'cloudy' but the inference is made that P input in bogs is very limited and may restrict N fixation and thus sites with high P availability may also be those with high rates of N fixation (and thus N concentration and, inferentially, N accumulation). As with all such analyses, the conclusions reached are dependent on the range of data collected and interdependence among some of the parameters that are introduced into correlations and regression models. As with many analyses of this nature, it raises more questions than it answers, which is fine.

It is well written and structured and delightfully succinct (6 pages without the references and tables/figures), which is justified by the data base. Perhaps this is what Letters are supposed to be.

I offer the following comments and suggestions, in a constructive manner, by line number:

32 somewhere you need to define what 'surface' means and the Abstract would be a logical place to start this, as N and P concentrations vary with depth in the peat and advancement in decomposition. By searching the Supplemental data, it appears that the depth of sampling is variable but generally within the top 25 cm. As changes in N content and C:N ratio can occur within this depth, some acknowledgment needs to be made that depth of sampling maybe a significant variable in the data analysis (but not incorporated). This is dealt with on line 108.

**We agree, and have defined the depths in the Abstract (line 31), and added the mean value to the text on line 111.**

62 the sequence of reference citation seems random - neither alphabetical nor chronological.

**We have corrected all the multiple citations so that they are chronological.**

69 if the long-term rate is 0.5 g, one must add any losses by denitrification (minimal) and lateral export as DIN and DON, mainly in the latter form. Fluvial N export may be 0.3-0.5 g (depending primarily on runoff and DOC concentration), so the N fixation rate required is more likely to be about 1 g. As far as I am aware, this is larger than most N fixation estimates in bogs (Vile et al. excepted).

**The text has been edited to make the different fluxes clearer. Lines 66-74. An additional reference to N fixation data has been added (Hemond).**

81 I am not an ornithologist, but what is bird strike? Bird shit? As bogs contain little vegetation that is readily edible (compared to marshes and swamps and probably fens), I imagine that this source is also small compared to other peatlands.

**Strike is indeed a euphemism for the more scatological term preferred by the Referee. Birds don't necessarily visit bogs to feed, they also roost. We have amended the text to also mention mammals, and used the word "activities" to cover all processes. Lines 80-81.**

125 Yes, P seems to be more variable than N.

**No change needed.**

126 two 'wides' is a sentence is a bit awkward and redundant.

**The referee is right, we have changed the text. Line 125.**

128 one has to 'visualize' NP ratios, as they are not represented in the Table. The mean is 34, with a range from 6 to 138. The tropical peatlands do have a higher ratio than elsewhere (eyeballing). I think more could be made of the NP ratio. Although perhaps not reaching universal consensus, it appears that a NP ratio of about 15 indicates a plant's co-dependence on N and P [see Güsewell, S. N:P ratios in terrestrial plants: variation and functional significance. New Phytol. 164, 243 (2004)]. A larger ratio implies P deficiency/limitation and the inverse for NP <15. Of your 277 peat samples, only 15 have a ratio <15: i.e. P limitation is dominant throughout. One wonders (although I accept the dangers in using ratios wildly) whether there any relationships between NP ratio (the relative abundance) and the climatic variables. 4 of the 15 (with the lowest NP ratios) are from Turetsky's Canadian continental sites, where aeolian deposition of P may be higher (could be dust but also frequent forest fires as a source of P)..

For what it is worth, our own data (Wang et al. 2015) for bogs in Ontario show NP ratios averaging 20 and 21 for 0-10 and 10-20 cm depths (with ranges of 10 to 67 and 10 to 58, respectively) - quite consistent with your results. In these samples (n = about 77) for both, there is a strong correlations between [N] and [P], with an r<sup>2</sup> of 0.27 and 0.25. For what it is worth, the equations are (N in %, P in ppm):

$$N = 0.012 P + 0.363 \text{ and } N = 0.016P + 0.313$$

Not sure how they compare with your relationships (maybe it is buried in supplementary material). Importantly, our profiles show an increase in the NP ratio with depth in the profiles, which again provides support that these systems are P limited (N is buried, P is pulled out - faster than C - during the decomposition processes). You suggest this on 216.

**We think it is somewhat tenuous to link peat NP ratios to plant ratios, since the plants will be resorbing nutrients at litterfall, altering the stoichiometry of the peat relative to the vegetation. But we agree that it is worthwhile mentioning the NP ratios (new text in line 124, Table 1), and how they vary with climatic variables (new text in lines 137-141; 189-192). NP ratios are now also provided in our Supplementary database (Table S1).**

**We don't think our results have anything to add to the very interesting and important topic of N & P concentrations and ratios as a function of depth.**

**In the Wang et al 2015 paper, data for individual sites are not given, only averages for all sites. The mean values for surface bogs fall within the data used here (%N ~ 1.0, %P ~ 0.05). We don't see any point in adding this information to our paper. The equations that Time Moore provides don't make sense to us: for %P = 0.05 then P in ppm is 500, and the equations then predict %N of around 6-9%, way too high.**

186 It may be true that there is little correlation between current N deposition and N concentration in surface peat (the top 25 cm generally represents the past 100-200 years), but there is evidence that current N deposition influences N concentration in the capitulum of Sphagnum, as shown by Bragazza et al. (Global Change Biology (2005) 11: 106-114) for European data (see also Limpens et al. 2011) supported by me for a wider range of sites and treatments. Thus, there may a temporal discord between the two measures, and I think this is implicitly recognized in the manuscript. The Sphagnum is where N will be fixed.

**We agree that this point should be clear, and have altered the text accordingly (lines 193-201), with an additional reference (Bragazza et al 2015).**

215 I realise that this is leading to a recognition that N fixation in bogs may be limited by P concentrations in the sites at which this process occurs - dominantly the Sphagnum mosses. I would be cautious about N fixation rates, which were based on a presumed ARA:N<sub>2</sub> fixation ratio of 3:1, which may be incorrect as Vile suggests. One might note that her high rates of N fixation are for Sphagnum in Alberta peatlands - not far from Sphagnum with high P concentration and low NP ratios from Turetsky. Finally, I think one must also recognize that N fixation may arise from methanotrophy in peatlands. This has been shown, in the surface layer, by Larmola et al. [Methanotrophy induces nitrogen fixation during peatland development. PNAS 111, 734 (2014)], with small rates at her bog site; but there the water table is low, so that methane oxidation will occur beneath the surface. Water table position has not been mentioned in the ms (I assume that the 0-25 cm depth is in the acrotelm at most sites), but it may be a way where MAP assumes some importance (more rain = higher water table?).

**The referee is right to question our reliance on the results of a single study (Vile et al) here, and so we have added two other references that report high N fixation rates in bogs. We do not think that we can enter into a critique of measured and published N fixation rates.**

**The methanotrophy point is interesting, but on consulting the Larmola et al paper in PNAS we found that it explicitly excludes bogs – the methanotrophy effect in their work only applied to the fen stage of peat development. We therefore do not think that mentioning the Larmola study as suggested by the referee is appropriate.**

226 I agree that nutrients need to be brought into C-centric models of peatlands, though incorporating dominantly organic forms of N and P may be challenging.

**We are trying to do such modelling ourselves, and Wu & Blodeau have published a C-N peat model, now included in the references.**

**Reviewer #3:**

This is a strong paper which would interest the readership of Biogeochemistry.

It would be good to see the scope broaden to maximise the relevance of the paper across fields.

**It is not at all obvious to us that citing and discussing the suggested additional references would improve the paper, rather it might reduce the focus, something very much to be avoided in a Letter.**

Suggestions below:

L85-89; L170; L222 The importance of phosphatase in organic P acquisition could be described Soil Biology & Biochemistry 31, 449-454, and in particular related to the wider response to nutrient / P constraints (Functional Ecology 19: 582-593.)

**We cannot see that a description and discussion of phosphatase is directly relevant to the current findings. Obviously it is part of the peat nutrient cycling system, but our results cannot say anything about this enzyme, and we prefer not to speculate. Really, this paper is about new facts and possible implications for the acquisition of elements by bogs, not a review of nutrient cycling in peats.**

L74 The section on the importance of ombrotrophic denitrification, should also refer to key papers on this topic (there are currently not references) e.g. Biogeochemistry 44, 151-162; Env. Sci. Technology, 31; 2438-2440.

**Denitrification is mentioned as a flux term in peat N balance, but that is all. It doesn't seem logical to cite "key papers".**

L89 Bearing in mind the importance of hydrological constraints in the data, it would be interesting to comment on the interactive impacts of soil moisture deficit and N abundance on constraint of decomposition through the enzymic latch (Proceedings of the National Academy of Sciences 103(51): 19386-19389; Nature. 409, 149.)

**We don't agree that this would be interesting, since decomposition of peat organic matter is not directly relevant to the acquisition of new N.**

Overall, a paper that could be published with minimal revision.

**We are of course very pleased that this reviewer thinks the paper is strong and requires minimal revision. However, we feel that the only constructive suggestions really are to add peripheral references to the Introduction, which we have striven to keep crisp and to the point and so we feel therefore that it would divert from this aim to add any additional references to the manuscript as it stands.**

Electronic Supplementary Material

[Click here to download Electronic Supplementary Material: Toberman et al\\_Table S1 Surface peat N & P database\\_REVISED.xlsx](#)

